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Morphological drivers of trophic cascades

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Worldwide, local anthropogenic extinctions have recently been reported to induce trophic cascades, defined as perturbations of top consumers that propagate along food chains down to primary producers. This focus on the effects of top-consumer extinction (i.e. of species presence) ignores potential cascading effects of the rapid morphological changes that may precede extinction. Here, we show in an experimental, three-level food chain including medaka fish, herbivorous zooplankton and unicellular algae that varying body length of a single fish from large (36.3 mm) to small (11.5 mm) induced a stronger trophic cascade than varying an average-sized (23.8 mm) fish from being present to absent. The strength of fish predation on zooplankton scaled quasi linearly (not with a power exponent) with fish body length and associated gape width, suggesting that the resultant trophic cascade was morphology (not metabolism)-dependent. The effect of fish body length was stronger on phyto- than on zooplankton, because large-sized fish had the unique ability to suppress large-sized herbivores, which in turn had high grazing capacities. Hence, our results show that consumer body size, by setting diet breadth, can both drive and magnify the strength of trophic cascades. In contrast, fish body shape had no significant effect on fish predatory performances when its allometric component (the effect of size on shape) was removed. In the wild, human-induced body downsizing of top consumers is widespread, and mitigating the resultant perturbations to ecosystem function and services will require a paradigm shift from preserving species presence towards preserving species size structure.

Biodiversity loss caused by human activities is not random, but mainly affects large vertebrates at the top of trophic networks (Duffy 2003). In the impacted populations, collapse is often preceded by rapid changes in phenotypic traits (Olsen et al. 2004, Palkovacs et al. 2012). Best documented is a rapid adaptive change towards smaller body sizes, as demonstrated in exploited populations of fish (Olsen et al. 2004, Edeline et al. 2007, Darimont et al. 2009), mammals (Coltman et al. 2003) and plants (Law and Salick 2005). At the ecosystem level, exploitation generates anthropogenic trophic cascades – ATCs – (Daskalov et al. 2007, Estes et al. 2011) which refer to indirect effects of perturbation to top consumers cascading along food chains down to primary producers (Carpenter and Kitchell 1993). In food chains with odd numbers of species (e.g. three-level food chains), top predators release primary producers from herbivore control, while in even food chains top predators favour the control of primary producers by herbivores (Oksanen et al. 1981).

Despite an early recognition of the importance of large-sized herbivores to the very existence of trophic cascades in

lakes (Carpenter and Kitchell 1993), trophic cascade ecology often remains centred around studying the effects of local top consumer extinction (i.e. presence versus absence), and ignores potential for trait-dependent effects (Bolker et al. 2003, Persson et al. 2003, Werner and Peacor 2003). For instance, ATCs remain generally ascribed to local extinctions of top consumers in all of the world major biomes (Estes et al. 2011). Similarly, experimental studies on trophic cascades almost exclusively vary presence versus absence of top predators (Hulot et al. 2014 but see Lazzaro et al. 2009), and ATCs in marine systems are ascribed to declining top-predator densities with no consideration for the role of size truncation resulting from size-biased mortality (Daskalov et al. 2007, Baum and Worm 2009, Estes et al. 2011).

However, traits are key drivers of predatory performances. In particular, metabolic rates scale with body mass to a power exponent such that attack rates increase and handling times decrease with an increasing body size, while in parallel gape width and diet breadth increase linearly (Peters 1983, Woodward et al. 2005, Barneche et al. 2014). Therefore,

there is a high potential for anthropogenic body downsizing per se to lower the strength of predator-prey links and to drive size-dependent ATCs (Persson et al. 2003).

Accordingly, Shackell et al. (2010) recently showed that body-downsizing caused by commercial fishing in populations of predatory fish induced an ATC in the Northwest Atlantic, despite that the biomass of predators remained constant, i.e. despite that predator numbers actually increased. Given that metabolism, including feeding rate, scales with body mass to a power exponent lower than one (often with a $3/4$ exponent, Peters 1983, Woodward et al. 2005, Barneche et al. 2014), increased numbers of smaller-sized predatory fish should result in an overall increased prey consumption (see Supplementary material Appendix 1 for a formal demonstration). In contrast, the observed trophic cascade in the northwest Atlantic shows a relaxed predation pressure from predatory fish (Shackell et al. 2010). Therefore, the ATC in northwest Atlantic was not density- or physiology (metabolism)-dependent, but was instead size (morphology)-dependent. This important observation, made on a large natural system, urgently calls for an improved understanding of the mechanisms driving size-dependent trophic cascades, and for a quantification of the relative contributions of consumer presence versus body size to the strength of trophic cascades. Unravelling these mechanisms and quantifying these relative contributions in an experimental setting was the first aim of our study.

The second aim of our study was to disentangle the relative contributions of length, shape and allometry (shape explained by length) in driving the size-dependency of trophic cascades. Many fish species present distinct body shapes associated with distinct trophic specializations called pelagic (limnetic) versus benthic morphs in lacustrine habitats (Robinson and Wilson 1994). Human predators may induce changes in body shape and the trophic morphology of impacted populations, and thus potentially alter resource use and drive shape-mediated ATCs (Haas et al. 2010, Franssen 2011). For instance, in the guppy *Poecilia reticulata* presence of large-sized predators (mimicking guppy exploitation by humans) induces subtle evolutionary changes in food acquisition strategies towards reduced consumption of algae and detritus and increased consumption of invertebrates (Zandonà et al. 2011). These changes in prey selectivity seem correlated with changes in guppy head shape (Palkovacs et al. 2011). However, body shape is generally linked to body length through an allometric component (Gould 1966), and failure to account for this allometry may erroneously ascribe to shape effects that are in fact size-dependent. Here, we explicitly separate the allometric and non-allometric components of shape to study their relative effects on predatory performances.

Specifically, we used an experimental, three-level food chain in which we varied in parallel both presence and body size of medaka fish *Oryzias latipes* (the top predator), and measured the resultant cascading effects on its herbivorous zooplankton prey and on unicellular algae grazed by zooplankton. We found that fish body size, but not shape without allometric component, drove the strength of the trophic cascade by setting the strength of predation experienced by large-sized herbivore, which in turn were the most efficient grazers.

Material and methods

Experimental fish

Medaka is a small freshwater fish from southeast Asia. Its small size (25–35 mm as adult), extreme resistance to environmental changes (temperature and salinity) and to manipulations make medaka an ideal model organism in biology (Kinoshita et al. 2009, Kirchmaier et al. 2015). In the wild, medaka inhabit stagnant and relatively shallow water (e.g. rice fields and irrigation canals). It feeds preferably on zooplankton (cladocerans and copepods) but also on annelids, larvae of Dipterans and even on macroalgae (filamentous algae or *Closterium* sp.) and benthic detritus when no other food is available (Terao 1985). Recently, it was shown that medaka from different populations of the Nagano Prefecture in Japan exhibit covariation between predatory behaviour (tendency to feed on surface versus benthic food items) and body shape (Iguchi and Kitano 2008), and that predation efficiency of adult medaka (Standard body length measured from the snout to the base of the caudal fin – SL – ranging from 26 to 36 mm) on *Daphnia pulex* may be dependent on *Daphnia* body size (Mano and Tanaka 2012). These studies suggest that medaka is a particularly good model predator to explore trait-dependent trophic interactions.

The medaka used for the present experiment (SL ranged from 11.5 to 36.3 mm, mean \pm SD = 23.8 ± 7.5 mm) descended from 100 wild medaka caught in Kiyoshu (Toyohashi, Aichi Prefecture, Japan) in 2011, and stocked in artificial, outdoor ponds at CEREEP Ecotron Île-de-France near Paris, France (<www.foljuif.ens.fr>). In the artificial ponds, medaka naturally feed on zooplankton and invertebrate communities and reproduce from May to August. Here, we used fish sampled from a pond in February, and acclimatized to laboratory conditions during one week prior to the experiments.

Behavioural assays

Experiments were conducted at $24 \pm 1^\circ\text{C}$ under artificial light conditions (12 h-light:12 h-dark cycle). We used 60 aquaria (vol = 1 l) containing initially a two-level food chain with phytoplankton (*Scenedesmus* spp.) and herbivorous zooplankton sampled in a nearby natural pond (95% *Daphnia pulex* – hereafter *Daphnia* – and 5% cyclopoid copepods present in both nauplius and copepodite stages). We generated pairs of identical aquaria ($n = 30$ pairs per experiment) using a plankton splitter, which separates a parent sample of live plankton in twin samples (see photography in Supplementary material Appendix 2 Fig. A1).

A pilot methodological experiment showed that the plankton splitter generated identical algal concentrations, *Daphnia* counts and copepod counts in the two aquaria of a given pair (see analyses in Supplementary material Appendix 2 and script in Appendix 9). Hence, by generating no within-pair (i.e. between-treatment) variance in counts but only between-pair (i.e. within-treatment or residual) variance, our splitting method was conservative regarding the mean estimate of treatment effect on counts, and could have only decreased (not increased) the ratio between explained deviance and residual deviance in our ANOVAs (i.e. no added risk of false positive in models 1–3, Table 1). Note that

Table 1. Summary of the statistical models used to test presence- versus size-dependent effect of medaka on the food chain. Coefficient values were estimated against the intercept reference levels that represent zooplankton counts (log scale) or phytoplankton concentration when fish was absent in experiment 1 (model 1–3), predation probability on the smallest size class of *Daphnia* by the smallest sized fish (logit scale, model 4), predation probability on nauplius larvae in experiment 1 (logit scale, model 5), or grazing probability on phytoplankton in experiment 1 when the smallest sized fish was present (logit scale, model 6). In model 3 and 6, main effects (a and b) are from a model without an interaction term (by definition, main effects are without the added effect of an interaction). Fish length unit is mm, *Daphnia* size class unit is mm². (:) indicates interaction between two predictors. (*) indicates overdispersed models, (¥) indicates that model 3 included residual heteroskedasticity as an exponential function (varExp within the gls function of the nlme library, Pinheiro et al. 2012) of fitted values v_i : $\text{var}(\epsilon_i) = \sigma^2 \exp(2\delta v_i)$ where ϵ_i = residuals and σ^2 = variance of the normally-distributed residuals, as estimated with no heteroskedastic structure (Pinheiro and Bates 2000).

Model	Trophic level	Response	Distribution	Link	Predictors	Coefficient value	Statistic value	p-value	Pseudo-R ²
1	Zooplankton	<i>Daphnia</i> count (n = 360)	Poisson*	log	intercept	3.80	t = 20.50	p < 0.001	0.44
					experiment 2	1.14	F = 38.04	p < 0.001	
					fish presence	-1.40	F = 53.35	p < 0.001	
					<i>Daphnia</i> size class	-6.70	F = 58.14	p < 0.001	
2	Zooplankton	copepods count (n = 240)	Poisson*	log	intercept	0.62	t = 3.84	p < 0.001	0.68
					experiment 2	2.06	F = 237.95	p < 0.001	
					fish presence	-1.26	F = 115.38	p < 0.001	
3	Phytoplankton	phytoplankton concentration (n = 1,200)	Gaussian¥	identity	intercept	75.67	t = 72.00	p < 0.001	0.80
					experiment 2 (a)	-70.77	F = 4631.00	p < 0.001	
					fish presence (b)	9.73	F = 328.00	p < 0.001	
					a:b	-9.35	F = 18.00	p < 0.001	
4	Zooplankton	predation probability on <i>Daphnia</i> (n = 180)	binomial*	logit	intercept	-4.51	t = -11.74	p < 0.001	0.75
					fish length	0.30	F = 379.99	p < 0.001	
					<i>Daphnia</i> size class	-11.75	F = 32.90	p < 0.001	
5	Zooplankton	predation probability on copepods (n = 120)	binomial*	logit	intercept	0.50	t = 1.11	0.269	0.07
					experiment 2	0.98	F = 5.14	0.020	
					copepodites stage	-0.63	F = 4.38	0.040	
6	Phytoplankton	grazing probability on phytoplankton (n = 600)	binomial*	logit	intercept	-2.69	t = -22.25	p < 0.001	0.50
					experiment 2 (a)	1.60	F = 489.60	p < 0.001	
					fish length (b)	0.06	F = 173.71	p < 0.001	
					a:b	0.05	F = 12.73	p < 0.001	

adding a fixed ‘aquarium-pair’ effect in models 1–3 (Table 1) did change the number of degrees of freedom, but did not affect model estimates or p-values in the F-tests.

Although counts were homogeneously split by our method, we found that copepodites and *Daphnia* tended to be slightly but significantly larger in the left branch of the splitter (Supplementary material Appendix 2). Hence, we spread this bias among the fish-present and fish-absent food chains (see below for fish addition) by alternating the position of the fishless aquarium between the left and right branches of the splitter. Therefore this bias could have only added noise to the data without any directional effect.

After 24 h fasting, a single medaka was added in one aquarium of each aquarium pair (n = 30 fish), thus generating a three-level food chain, and the fish was allowed to forage on zooplankton during 24 h. Then, the fish was removed, anaesthetized with tricaine methanesulfonate (3×10^{-4} g l⁻¹), photographed from lateral and dorsal views (for subsequent morphometric measurements), and released. Zooplankton in each aquarium was fixed in ethanol (96%) for subsequent enumeration and size measurement using the ZooScan (Gorsky et al. 2010). This device takes a high-resolution picture of the zooplankton sample, extracts each particle using the Zooprocess V.7.091 software, measures each particle size, and assigns it to a taxonomic category based on a “learning set” provided by the user using the Plankton Identifier V.1.261 software. We further complemented this automatic sorting by a manual check. Zooplankton body length (longest axis) ranged from 2.0×10^{-2} to 2.0 mm in *Daphnia* (median = 2.4×10^{-1}), from 5.7×10^{-2} to 1.7×10^{-1} mm in

nauplius larvae (median = 6.8×10^{-2}), and from 1.1×10^{-1} to 5.9×10^{-1} mm in copepodites (median = 1.2×10^{-1}). In our analyses (below), we assigned *Daphnia* into three size classes based on the log of the scanned body surface area, which provided a more accurate proxy for body size than the longest axis (central values of size classes: 3.1×10^{-1} , 4.7×10^{-2} and 7.2×10^{-3} mm²). We assigned copepods into two stage classes (nauplius and copepodite stages, respective central values : 3.0×10^{-3} and 7.2×10^{-3} mm²).

We measured phytoplankton concentration (in µg l⁻¹ chlorophyll a) in each aquarium using a spectrofluorimetric probe BBE, which measures the fluorescence emitted by the chlorophyll pigment content in response to excitement by light emitted by diodes. We recorded 10 measurements for each sample.

This whole experiment including behavioural tests, plankton and fish morphometric measurements was repeated twice (hereafter experiment 1 or 2), yielding a total of 120 aquaria and 60 fish.

Medaka morphometric measurements

We measured each fish body length (n = 60 SL measurements) to the nearest 10⁻¹ mm using ImageJ V.1.49m software (Abramoff et al. 2004).

In order to quantify medaka body shape (Adams et al. 2004), we placed both landmarks and semi-landmarks on each medaka picture (Mitteroecker and Gunz 2009). Landmarks are defined as fixed homologous points in terms of anatomical structures, and semi-landmarks are sliding points

approximately placed onto medaka contour by the observer and which final position is determined by minimizing the Procrustes distance criterion (Bookstein 1996). We digitized 21 points (8 landmarks and 13 semi-landmarks) onto the dorsal head shape (see photography in Supplementary material Appendix 3 Fig. A2_A) and 19 points (11 landmarks and 8 semi-landmarks) onto the lateral body shape of medaka (Supplementary material Appendix 3 Fig. A2_B) using tpsDig2 V.2.16 software (Rohlf 2010). We used the “geomorph” library (Adams et al. 2004) in R V.3.1.0 software (<www.r-project.org>) to perform generalized Procrustes analysis – GPA – (gpagen function of ‘geomorph’ library) and to remove non-shape variation (the effects of rotation, translation and the isometric effects of size). The GPA superimposes all dot patterns and performs three operations: a centring configuration, a size normalization (scaling), and a rotation calculated in order to minimize the sum of squared deviations between the coordinates of a same point (Rohlf and Slice 1990). The GPA further provides a consensus (the average fish body shape), the centroid size of each individual (the size before scaling), and the Procrustes shape coordinates (containing information on the superposed body shape).

In a pilot methodological experiment, we showed that this method allowed us to accurately discriminate individual medaka according to their shape, and was thus robust to the noise introduced by the digitizing procedure (see analyses in Supplementary material Appendix 4 and scripts in Appendix 9 and 11).

Statistical analyses

Presence - versus size-dependency of the trophic cascade

We modelled the general effects of fish presence on zooplankton counts and on phytoplankton concentrations in analyses of variance, in which each observations were made at the level of each single aquarium (models 1–3 in Table 1). Model 1 included 360 observations (120 aquaria \times 3 *Daphnia* size classes), model 2 included 240 observations (120 aquaria \times 2 copepod stage classes), and model 3 included 1200 observations (120 aquaria \times 10 repeated measurements).

In contrast, modelling the effect of medaka body size on predation or grazing probability required using binomial GLM regressions (models 5–6 in Table 1) in which an observation was provided by an aquarium pair ($n = 60$ aquarium pairs), and in which response was defined as:

$$p_{zoo} = \frac{A_z - P_z}{A_z} \text{ or } p_{phyt} = \frac{P_p - A_p}{P_p}$$

Where A_z = zooplankton count in a given size class (*Daphnia*, 60 aquarium pairs \times 3 sizes classes = 180 observations) or development stage (copepods, 60 aquarium pairs \times 2 stage classes = 120 observations) in the fish-absent aquarium of an aquarium pair, A_p = phytoplankton concentration in the fish-absent aquarium of an aquarium pair (60 aquarium pairs \times 10 repeated measures = 600 observations), P_z and P_p = same variables in the fish-present aquarium of an aquarium pair. Hence, p_{zoo} = probability for zooplankton to be eaten by fish, p_{phyt} = probability for phytoplankton to be grazed by zooplankton when the fish is removed (i.e. in the fish-absent aquarium compared to the fish-present aquarium).

In addition to the effects of fish presence or body length on the food chain, we tested in our models for the effects of replicating the experiment (experiment 1 or 2), and of zooplankton size class (real scale, coded as a continuous variable) or development stage (nauplius larvae versus copepodite stage, coded as a categorical variable), alone or in interaction with fish presence or fish body length. Because predictors were different for *Daphnia* (body size class) and copepods (development stage), we analysed their response to predation in two separate models. In our models, we also tested the effect of fish body mass and found it to be a weaker predictor of medaka predation efficiency on *Daphnia* (6% less deviance explained) than fish body length. Finally, we also found gape width to be a slightly better predictor of medaka predation efficiency on *Daphnia* (0.8% more deviance explained) than fish body length. However, we chose to keep body length in our analyses because it is a more universal morphological metric.

In selecting our models, we kept or discarded each effect based on AIC change when the effect was removed from the model. Only changes larger than two points in AIC – or qAIC for overdispersed models (Bolker 2014a), – were considered (Burnham and Anderson 2002). If AIC or qAIC change was not significant, the effect was discarded to keep the most parsimonious model. The selected models are presented in Table 1. We computed goodness of fit of the selected models as the pseudo- R^2 (Table 1), i.e. as the deviance ratio between the selected model and the null model which included an intercept as the only effect. All statistical analyses were performed in R 3.1.0 software (<www.r-project.org>) with the ‘nlme’ (restricted maximum likelihood; Pinheiro et al. 2012) and ‘bbmle’ (maximum likelihood; Bolker 2014b) libraries (see scripts in Supplementary material Appendix 5 and 10).

Zooplankton counts were overdispersed (median count = 3, mean count = 23), and we thus explored two different classes of overdispersed count models based on estimating a multiplicative dispersion parameter (quasipoisson in the glm function of R) or on a negative binomial distribution (zero-inflated mixture models had convergence problems with the full set of predictors). We found the model selection procedure to yield similar results with both model classes, and we chose to keep the quasipoisson model for its wide use and its standard implementation in the glm function of R.

Computation of effect sizes

We compared the relative contributions of medaka presence vs. body length to the strength of the trophic cascade using effects sizes on zoo- and phytoplankton predation probability, computed from overdispersed binomial GLMs similar to models 4–6 (Table 1, see also script in Supplementary material Appendix 6).

We computed the effect size of fish presence as μ = mean probability for prey to be eaten, as predicted from an intercept-only GLM fitted to the data using maximum likelihood. This was technically equivalent to comparing the effect of a 0.0 mm fish (i.e. medaka absence) with the effect of a medium-sized (SL = 23.8 mm) fish, and could be written as $\mu = \overline{P_{23.8}} - \overline{P_0}$, where $\overline{P_{23.8}}$ = mean probability for prey to be eaten when the fish was present, and $\overline{P_0}$ = mean probability for prey to be eaten when no fish was present.

We computed the effect size of fish body length as $s = \overline{P}_{36.3} - \overline{P}_{11.5}$, where $\overline{P}_{36.3}$ = mean probability for prey to be eaten when the largest fish in the population (SL = 36.3 mm) was present, and $\overline{P}_{11.5}$ = mean probability for prey to be eaten when the smallest fish in the population (SL = 11.5 mm) was present. Both $\overline{P}_{36.3}$ and $\overline{P}_{11.5}$ were predicted from an overdispersed GLM fitted to the data by maximum likelihood, and which included an intercept plus fish body length as predictors. Thus, dividing μ by 23.8 and s by $36.3 - 11.5 = 24.8$ would yield the slopes of the linear relationships between predation probability on plankton and medaka body length. We estimated mean and standard deviation for both μ and s in a bootstrap procedure in which we randomly sampled half of the observations (with replacement), fitted the models and recorded parameters, repeated 1000 times.

Additionally, we complemented this bootstrap procedure by a Markov chain Monte Carlo (MCMC) approach in JAGS ver. 3.4 (Plummer 2003). Specifically, we fitted an overdispersed binomial model to each full count dataset (*Daphnia*, copepods and algae). Following Bayesian standards, we modelled overdispersion as an individual-observation random effect (Kéry 2010). The models included an intercept μ and a slope for the effect of medaka body length (standardized to 0 mean and unity standard deviation), from which we computed s as defined above. Additionally, we computed the μ/s ratio as part of the MCMC chains.

We ran three independent chains of 50 000 iterations each with a burn-in period of 40 000 iterations. We thinned the three chains every 5th iteration, yielding a posterior of 6000 samples for each parameter. Uninformative priors for regression parameters were defined as normal distributions with zero mean and 100 standard deviation, and for variance parameters as a uniform distribution between 0 and 10. Convergence was assessed using the Gelman–Rubin statistic (Gelman and Rubin 1992). The MCMC p-value for the μ/s ratio was defined as twice the proportion of the posterior which was larger than one when the posterior mode was less than one, or as twice the proportion of the posterior which was less than one when the posterior mode was larger than one. This test was equivalent to a bilateral t -test in which H_0 was $\mu = s$ and H_1 was $\mu \neq s$. See script in Supplementary material Appendix 7 for more details about our MCMC approach.

Links between medaka body shape and predatory performance

We explored how medaka body shape, allometry-free shape (i.e. size-independent body shape), and allometric shape (i.e. size-dependent body shape) influenced predatory performance using partial least square regressions (two blocks partial least squares – 2B-PLS – Rohlf and Corti 2000). 2B-PLS (two.b.pls function of ‘geomorph’ library) diagonalizes one of the two matrices of covariance between the two data blocks and produces pairs of PLS orthogonal axes (PLS axes) that maximize the covariation between the two data blocks. The result of a 2B-PLS is thus similar to the output of a principal component analysis (redefinition of orthogonal axes in the multivariate morphological space), except that the PLS define morphological axes that are the most correlated to predation variables. In all three 2B-PLS regressions (Table 2, one 2B-PLS per fish view and one with the combined views), predation variables constituted the multivariate response, i.e. the five probabilities of zooplankton to be consumed by medaka (three *Daphnia* size classes + two copepod stages).

We extracted allometry-free and allometric shapes from multivariate analysis of variance (MANOVA, (Monteiro 1999); manova function of R) between body shape and log-transformed centroid size (Table 2, one MANOVA per fish view; We previously checked with a MANCOVA that fish body shape did not vary with experiment and/or sex. See script in Supplementary material Appendix 8). Residuals from the MANOVA provided allometry-free residuals that correspond to the size-independent body shape coordinates (i.e. allometry-free shape), while predicted shape values from the MANOVA provided the size-dependent body shape coordinates (i.e. allometric shape). We further computed the proportion of variance in shape explained by allometry as A/B , where A = allometric shape variance, and B = total shape variance, as predicted by the MANOVA.

Morphometric analyses were performed with the ‘geomorph’ library for R (Adams and Otárola-Castillo 2013), and allometric shape change were depicted using the tpsRegr V1.28 software (Rohlf 2003).

Data deposition

Data available from the Dryad Digital Repository: <<http://dx.doi.org/10.5061/dryad.6m1m6>> (Renneville et al. 2015).

Table 2. Summary of morphometric analyses. Correlations in 2B-PLS analyses (Cor values) are between the fish shape and predatory behaviour. Allometry-free shape are residuals from the regression of log-transformed centroid size on shape and allometric shape are the predicted values (see Methods for more details).

Analysis type	Response	Predictors	Fish view					
			Lateral view		Dorsal view		Lateral + dorsal views	
MANOVA	body shape	log-transformed centroid size	Pillai = 0.95	p < 0.001	Pillai = 0.89	p < 0.001		
2B-PLS shape	zooplankton proportions	body shape	Cor = 0.60	p < 0.001	Cor = 0.69	p < 0.001	Cor = 0.72	p < 0.001
2B-PLS allometry free shape	zooplankton proportions	allometry free shape	Cor = 0.30	p = 0.894	Cor = 0.45	p = 0.297	Cor = 0.48	p = 0.498
2B-PLS allometric shape	zooplankton proportions	allometric shape	Cor = 0.77	p < 0.001	Cor = 0.80	p < 0.001	Cor = 0.80	p < 0.001

Results

Presence – versus size-dependency of the trophic cascade

Presence-dependent effects

Presence of an average fish significantly decreased zooplankton abundance (Table 1, model 1 and 2) and significantly increased phytoplankton concentration (Table 1, model 3), indicating that medaka induced a trophic cascade in our aquaria.

Small-sized *Daphnia* were significantly more numerous than large-sized *Daphnia* (Table 1, model 1), while there was no abundance difference between the two copepod development stages (Table 1, ‘copepod stage’ discarded from model 2). There was no significant interaction between medaka presence and *Daphnia* size class or copepod development stage (discarded from model 1 and 2), indicating that an average medaka had no significant foraging preference among zooplankton sizes or stages (but see size-dependent effects below).

The significant effect of experiment on zooplankton count and phytoplankton concentration (Table 1, model 1–3) shows that zooplankton was more abundant and phytoplankton less concentrated during the second experiment. Finally, we found a significant interaction between fish presence and experiment on phytoplankton concentration (Table 1, model 3), indicating that fish presence increased less sharply phytoplankton concentration during the second experiment.

Size-dependent effects

The probability for *Daphnia* to be eaten by medaka significantly increased with medaka body length (Table 1, model 4; Fig. 1) and, in turn, the probability for phytoplankton to

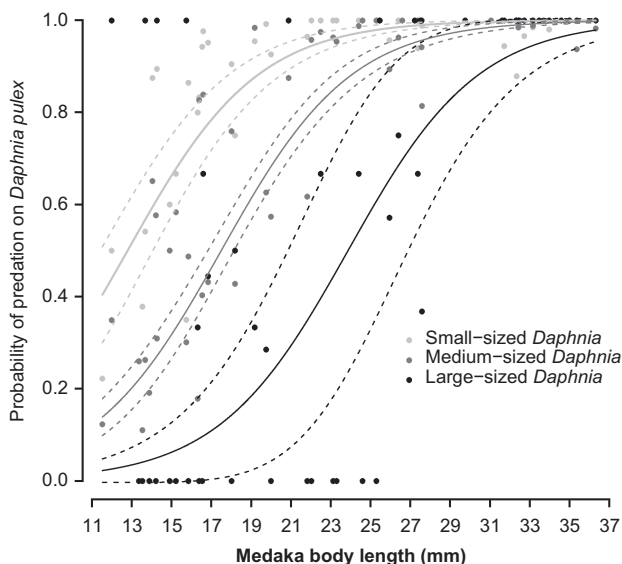


Figure 1. Size-dependent predation probability. Probability for *Daphnia* to be eaten by medaka as a function of medaka body length. Dots correspond to the raw data and lines correspond to model predictions with 95% confidence intervals (as predicted from model 4, Table 1).

be grazed by *Daphnia* significantly increased when a larger fish was removed (Table 1, model 6; Fig. 2). Therefore, medaka body length significantly influenced the strength of the trophic cascade induced by medaka. In contrast, medaka body length did not influence the probability for copepods (either copepodite or nauplius stage) to be eaten by medaka (‘fish length’ discarded from model 5 in Table 1), showing that the size-dependent trophic cascade was mediated by *Daphnia* only.

Small-sized *Daphnia* were significantly more consumed than large-sized *Daphnia* (Table 1, model 4), and the largest *Daphnia* were predated only by large medaka (Fig. 1). However, we did not find any significant interaction between *Daphnia* size and medaka length (discarded from model 4 in Table 1; Fig. 1), indicating that small and large medaka had the same preferences in terms of prey size.

Finally, we found that the probability for phytoplankton to be grazed by zooplankton overall increased during the second experiment, as compared with the first experiment. This effect was due to higher zooplankton numbers during the second experiment.

Comparison of effect sizes

The bootstrap suggested that the effect of varying medaka presence on zooplankton was 95% of the effect of varying medaka body length (Fig. 3). Specifically, an average-sized medaka increased predation probability on *Daphnia* by 75% (mean \pm 5.2 10^{-2} SD) compared to a fish-absent aquarium, while a large-sized medaka increased this probability by 79% (mean \pm 5.8 10^{-2} SD) compared to a small-sized medaka. In line with the bootstrap procedure, the MCMC approach suggested that the effect size of varying medaka presence on *Daphnia* was 98% (\pm 1.4 10^{-4} SD) of the effect

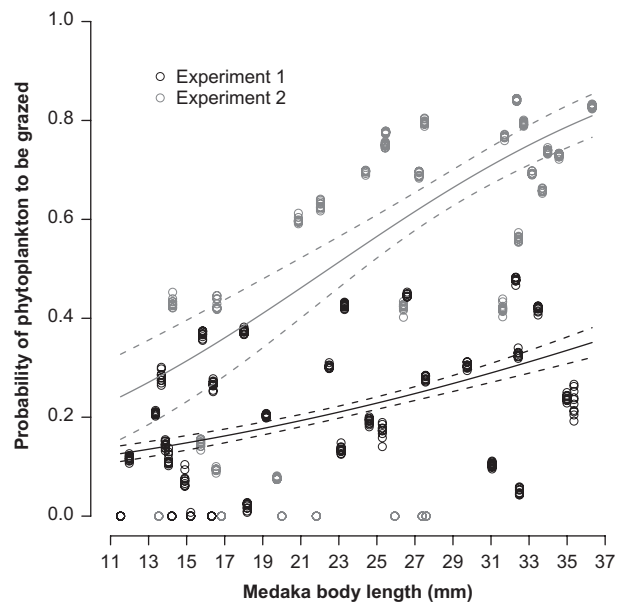


Figure 2. Size-dependent cascading effect. Probability for phytoplankton to be grazed by zooplankton when medaka is removed from an aquarium (i.e. change from fish presence to fish absence), as a function of medaka body length (see Methods for more details). Dots represent the raw data, and lines represent predictions with 95% confidence intervals (as predicted from model 6, Table 1).

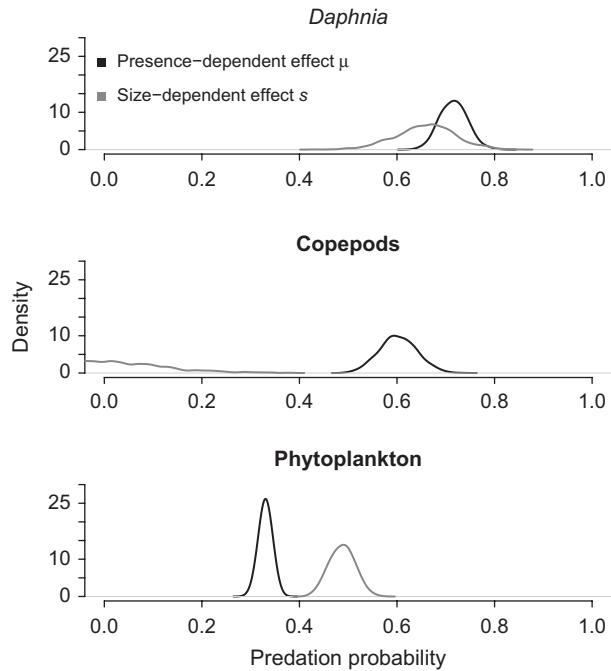


Figure 3. Density probability of effect sizes as estimated by bootstrap. The effect size of fish presence (presence-dependent effects μ , black line) indicates the mean change in probability for being eaten or grazed when an average medaka was added in a fishless aquarium (*Daphnia* and copepods) or when the average medaka was removed and aquarium returned to a fishless state (phytoplankton). The effect size of fish body length (size-dependent effects s , grey line) indicates the mean change in probability for being eaten or grazed when medaka body length was changed from minimum to maximum (*Daphnia* and copepods) or from maximum to minimum (phytoplankton). Predation probabilities were estimated 1000 times in a bootstrap sampling procedure (see text for more details). Densities were estimated using a Gaussian smoothing kernel with a standard deviation of 0.01.

of varying medaka body length. The MCMC p-value was 0.01, indicating that this difference was significant.

In contrast, the effect of medaka on copepods was almost only presence-dependent (Fig. 3). Specifically, an average-sized medaka increased predation probability on copepods (of either stage) by 73% (mean \pm 4.1 10^{-2} SD) compared to a fish-absent aquarium, while a large-sized medaka had almost no further effect (mean = 0.04 % \pm 1.2 10^{-1} SD) compared to a small-sized medaka. The MCMC approach confirmed that the effect of medaka body length on copepod counts was not significantly different from 0.

The effect of varying medaka presence on phytoplankton in the bootstrap was 77% of the effect of varying medaka body length (Fig. 3). Specifically, removal of an average-sized medaka increased grazing probability by 27% (mean \pm 1.0 $\times 10^{-2}$ SD) compared to a fish-absent aquarium, while removal of a large-sized medaka increased this probability by 35% (mean \pm 2.8 $\times 10^{-2}$ SD) compared to removal of a small-sized medaka. The MCMC approach suggested that the effect size of varying medaka presence was 38% (\pm 2.7 $\times 10^{-2}$ SD, MCMC p-value < 0.001) of the effect of varying medaka body length, i.e. suggested a stronger effect of size-dependency than the bootstrap procedure.

Links between medaka body shape and predatory performance

We found that the allometry explained 11% of variance in dorsal shape and 20% of variance in lateral shape. Allometric component visualization showed that large-sized medaka had a proportionally larger and more anterior mouth, a shorter skull and a stockier body than small-sized medaka (Fig. 4).

2B-PLS regressions showed a high correlation between body shape and fish predatory behaviour (Table 2, 2B-PLS shape), thus supporting results from the GLM-based analysis showing a significant effect of body length on medaka predatory performance and on the resultant strength of the trophic cascade (Table 1, model 4 and 6). However, we found no significant correlation between allometry-free shape and fish predatory performance (Table 2, 2B-PLS allometry-free shape), indicating that body shape after accounting for the allometric component did not significantly influence medaka predatory performances in our experiment. Accordingly, we logically found a significant effect of allometric shape on medaka predatory behaviour (Table 2, 2B-PLS allometric shape).

Discussion

As expected, we found that medaka presence induced a trophic cascade in which zooplankton was suppressed and unicellular algae were released from grazing by zooplankton. However, in contrast with the dominant focus on presence-dependent trophic cascades, we found that the strength of this trophic cascade was more size- than presence-dependent.

In our experiments, the increase in predation probability on *Daphnia* induced by varying medaka body length from small to large was 105% of the increase induced by varying medaka from absent to present. This effect-size difference parallels body-length difference between medaka being large or small (36.3–11.5 = 24.8 mm) and medaka being present or absent (23.8–0 = 23.8 mm), since $24.8/23.8 \times 100 \approx 105\%$. In other words, the slope of the relationship between predation strength and medaka body length was identical when varying independently body length from 0 to 23.8mm or from 11.5 to 36.3 mm. Therefore, medaka predation probability on *Daphnia* presumably scaled quasi-linearly with medaka body length as would be expected from morphology-dependent feeding rates, and not with a power exponent as would be expected from metabolism-dependent feeding rates (Peters 1983, Barneche et al. 2014). This result is in line with the results of Shackell et al. (2010), who found in the northwest Atlantic that fishing-induced body down-sizing of predatory fish drove a trophic cascade reflecting a relaxed strength of predation despite that their total number actually increased (see the introduction). Therefore, both experimental and empirical data point to morphological (but not metabolic) constraints on predator–prey interactions as prominent drivers of the strength of trophic cascades.

Linear size scaling of medaka predation strength on *Daphnia* was likely mediated by gape width and resultant diet breadth. Accordingly, diet breadth scales linearly with body size in both terrestrial (Sinclair et al. 2003) and aquatic habitats (Claessen et al. 2002, Woodward et al. 2005).

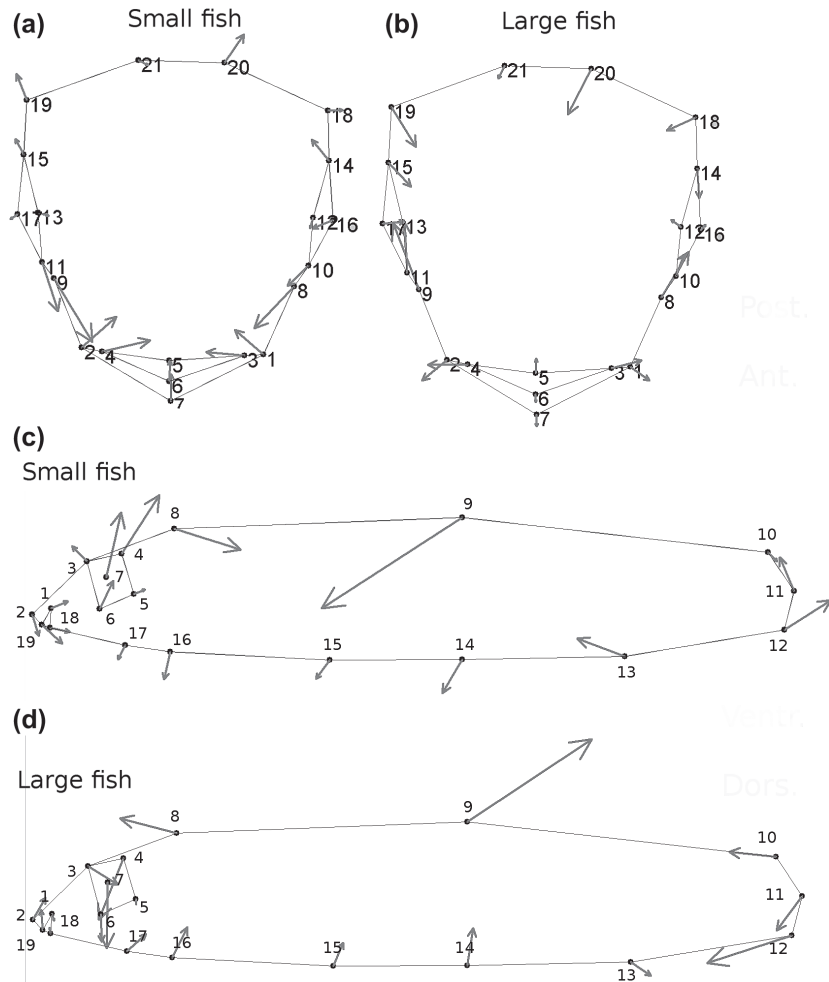


Figure 4. Medaka allometric shape. Shape variation between small and large medaka. The solid line represents the shape of an average fish. Vectors show the change from the average to the smallest specimen (a and c) and from the average to the largest specimen (c and d), for dorsal view (a and b; mouth down oriented; see also Supplementary material Appendix 3 Fig. A2) and lateral view (c and d; back down oriented; see also Supplementary material Appendix 3 Fig. A2). Differences (arrow length) are tenfold exaggerated to facilitate deformation visualization.

Hence, as medaka diet breadth was expanding in parallel with increasing medaka body length and gape width, a larger proportion of the *Daphnia* population became available for predation and the medaka-*Daphnia* predator-prey link was strengthening. Accordingly, gape width was a slightly better predictor than body length for medaka predation efficiency on large-sized *Daphnia*. In contrast, copepods, due to their small body size, were apparently included into the prey size range of both small and large medaka, such that the strength of the medaka-copepod, predator-prey link was not size-dependent. Therefore, our results, together with the study of Shackell et al. (2010), suggest that size-dependent diet breadth might play a central role in mediating the strength of predator-prey interactions, and from there in driving the strength of trophic cascades.

Importantly, comparison of effect sizes of medaka on zoo- and phytoplankton with both the bootstrap and MCMC approaches reveal that the contribution of size-dependency to the strength of the trophic cascade was magnified as the cascade was moving down along the food chain. This magnification apparently operated because large-sized medaka, in contrast with small and medium-sized medaka,

were capable of preying upon large *Daphnia* (Fig. 1), which in turn have the highest filtering capacities (Burns 1969). This result is in line with theoretical models predicting that the strength of trophic cascades is stronger in presence of a small plant/herbivore body size ratio, such that large-sized herbivores are highly efficient grazers and potent magnifiers of the strength of trophic cascades (Shurin and Seabloom 2005, DeLong et al. 2015). Accordingly, herbivorous *Daphnia* are one or two orders of magnitude larger than their unicellular algal food and are, as also underscored by our results, potent magnifiers of the strength of trophic cascades (Carpenter and Kitchell 1993). In grasslands, large mammalian herbivores are also orders of magnitude larger than their plant food, and they have indeed been shown to be central for the control of plant biomass (Shurin and Seabloom 2005, Wilmers et al. 2006, Holdo et al. 2009). Hence, we believe that our experimental results are qualitatively valid for a wide range of ecosystems, i.e. that size-dependent trophic cascades and their potential magnification at the herbivore-plant interface should systematically be considered in freshwater, marine or terrestrial biomes.

Beyond body length, we also explored whether medaka body shape influenced predatory performances and drove a

shape-mediated trophic cascade. We found that body shape including allometry was significantly linked with medaka predatory performances. However, body shape without its allometric component did not influence medaka predatory performances, highlighting the need to correctly account for the effect of body length on shape before testing for relationships between body shape and behaviour. This negative result in our experiment might be due to 1) an effective absence of any significant covariation between allometry-free shape and predatory behaviour at the within-population level, and/or 2) to the fact that we used only pelagic prey (plankton), while specialization in fish is often associated with the contrasted consumption of pelagic versus benthic prey (Robinson and Wilson 1994).

In human-perturbed ecosystems, the strength of ATCs is likely to be modulated by the amplitude of body downsizing experienced by top predators. The amplitude tested in our experiment (from 36.3 to 11.5 mm) is modest compared to anthropogenic changes observed in the wild. For instance, in the western Scotian Shelf ecosystem (northwest Atlantic) where fishing induced an ATC, planktivorous fish shrank from ca 35 to 30 cm, and benthivorous fish from ca 58 to 42 cm between 1970 and 2008 (Shackell et al. 2010). Often, however, body downsizing is more severe. In the southern Gulf of St Lawrence cod *Gadus morhua* mean length at maturity decreased from 65 to 40 cm in females and from 55 to 35 cm in males (Swain 2011). On longer time scales, decrease in mean body length may be even larger. Archaeological records of skeletal parts suggest that mean body length of cod in the Gulf of Maine (USA) has decreased from ca 100 to 30 cm due to overfishing (Jackson et al. 2001). In large part, such fishing-induced body downsizing reflects evolutionary changes (Coltman et al. 2003, Olsen et al. 2004, Edeline et al. 2007), and thus both trait change and resultant ATCs might be slow to reverse (Law 2000).

The size-dependency and potential poor reversibility of ATCs calls for a reassessment of conservation measures for biodiversity. Currently, management plans are focused on preserving minimum population numbers to avoid the negative effects of demographic stochasticity on population persistence, i.e. focus on preserving population presence. This classical conservationist perspective should be coupled with a preservation of population size structure, which would limit the strength of ATCs and maintain ecosystem functioning. Importantly, a functional, size-based management also meets conservationist objectives because large-sized individuals often have an increased fecundity and are thus more likely to contribute to increasing population numbers and to relaxing the effects of demographic stochasticity (Berkeley et al. 2004). Additionally, through their elevated egg production, large-sized females are also key to maintaining strong density-dependent regulation and dampen the effects of environmental stochasticity on population dynamics (Hidalgo et al. 2011, Botsford et al. 2014). Finally, preserving population size structure may protect from predator-prey reversals, in which anthropogenic body-downsizing results in downgraded trophic rank and increased predation mortality for the impacted population (Fauchald 2010). Taken together, these results suggest that integrating size-structure preservation into management plans would improve our ability to maintain biodiversity and restore ecosystems.

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Supplementary material (available online as Appendix oik-02877 at <www.oikosjournal.org/appendix/oik-02788>). complementary analyses: Appendix 1–4. Scripts: Appendix 5–11.